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Adsorptive Removal of Manganese (II) from Aqueous Solution using Graphene Oxide: A Kinetics and Thermodynamics Study

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ABSTRACT

An excess amount of manganese is known to affect neurological toxicity. This study focused on adsorptive removal of Mn(II) using graphene oxide, which was carried out under optimum conditions after varying some experimental effects including pH, an incubation time and an initial concentration. It was demonstrated that Mn(II) adsorption follows both Langmuir and Freundlich isotherms. The maximum adsorption capacity for Mn(II) removal from an aqueous solution was 41.67 mg/g at pH 5 and their adsorption state was completed within 30 min. The adsorption kinetics was in accordance with pseudo second-order kinetic model. In thermodynamics diversion, changes in free energy, enthalpy and entropy were also studied. Overall, the adsorption process was exothermic and spontaneous in nature.

> Keywords: Graphene oxide; Manganese; Adsorption isotherm; Thermodynamics; Kinetics

INTRODUCTION

Manganese (Mn) is an abundant element in the Earth's crust and its presence in water is a result of leaching process.¹ Although Mn is an essential nutrient element in living systems, an excess of Mn can result in toxic neurological effects, and there are many neurotoxic effects that cause a series of symptoms.²⁻⁴ Therefore, it is necessary to remove Mn(II) contaminants particularly from wastewater before it is released back to the environment. The adsorption process is one of the most promising for removal of metal ions. However, several carbonbased adsorbents for metal removal from wastewater are still limited in their use because of low efficiency and adsorption capacity. In order to improve capacity and selectivity of the adsorption, new adsorbents have been developed. Among these, graphene oxide (GO) is increasingly used as a new choice for carbon-based adsorbents. There are many applications of GO for removal from water samples of toxic metals and dye pollutants such as Hg(II),⁵ Pb(II),⁶ As(V),⁷ Cd(II),⁸ Cr(VI),⁹ Co(II),¹⁰ malachite green¹¹ and alizarin red S¹². However, there have not been reports of its use on the removal of Mn(II) from aqueous solution.

In this study, GO was applied to remove Mn(II) from aqueous solution. Potential use of GO for Mn(II) removal in a batch adsorption study was compared with that of other carbon-based adsorbents. The effects of pH, an incubation time and an initial concentration of Mn(II) were optimized. Langmuir and Freundlich isotherms were also investigated in detail to fit its adsorption model. Both kinetics and thermodynamics aspects were also investigated to find out the adsorption behaviour for Mn(II) from aqueous solution.

MATERIALS AND METHODS

Preparation of graphene oxide

In this work, graphene oxide was prepared from commercial graphite following a protocol described in our previous work.¹³ Briefly, 3.0 g of graphite, 300 mL of 98% H₂SO₄, 23.0 g of KMnO₄ and 3.0 g of NaNO₃ were mixed under cold condition (0 °C) in an ice bath. After a 30-min stirring, a slow addition of 200 mL of deionized water to the mixture followed. The mixture was subject to another 30 min of stirring and then heated to 80 °C for 6 h before an addition of 40 mL of 30% H₂O₂. We used deionized water to wash the mixture several times until the filtrate pH reached neutral at which point lyophilized precipitates could be obtained.

Adsorption experiment

The adsorption experiment of Mn(II) was performed using batch technique. Graphene oxide (0.02 g) was accurately weighed into a 125 mL conical flask followed by an addition of 25 mL of Mn(II) solution (10 mg/L). pH of the solution was adjusted. At room temperature (30 °C) and at approximately 200 rpm, the solution was subjected to shaking by an orbital shaker. To determine the optimum conditions for maximum adsorption capacity (q_m) of GO, various experimental parameters were considered including

pH of solution (1-6), an incubation time (1-50 min), and an initial concentration of Mn(II) (2-30 mg/L). After a period of shaking, we subjected the mixed solution to 5 min of centrifugation to separate the absorbent. AAS was used to determine Mn(II) in the supernatant solution. We conducted our all of our experiments in triplicate with the same experimental conditions. An equation below was used to determine the adsorption capacity (q_e , mg/g) of Mn(II) at an equilibrium state,

$$q_e = \frac{V(C_o - C_e)}{m} \qquad \dots (1)$$

where, C_o , C_e , V and m are the initial Mn(II) concentration (mg/L) of the solution, Mn(II) concentration at the equilibrium state, the solution volume (L) and the adsorbent mass (g), respectively. To describe the equilibrium nature of Mn(II) adsorption onto the adsorbent, Langmuir and Freundlich adsorption models were also studied.

RESULTS AND DISCUSSION

Optimum conditions for Mn(II) removal

Fig. 1(a) shows the effect of pH on the adsorption capacity. The pH value was adjusted by acetate buffer solution (pH 1-5) and phosphate buffer solution (pH 6). The results showed that the adsorption capacity of Mn(II) dramatically increased with the solution at pH 1-4 and was constant in the pH range of 5-6. The maximum adsorption capacity was at pH 5 (3.14 mg/g). The pH value at the point of zero charge (pH_{PZC}) of GO has been previously reported to be 4.111. At pH lower than pH_{PZC}, the adsorption capacities of GO were lower due to the fact that under acidic solution, the surface GO surface was positively charged and did not favor an uptake of Mn(II). Whereas, at pH higher than pH_{PZC}, the adsorbents' surface became negatively charged facilitating the uptake of Mn(II). As a result, we found that we had the maximum adsorption capacity of GO at pH higher than 5.

Table 1: Thermodynamic data of Mn(II) adsorption using GO as an adsorbent

∆H⁰ (kJ/mol)	ΔS° (kJ/mol.K)	$\Delta {f G}^{\scriptscriptstyle 0}$ (kJ/mol)				R ²
		303 K	313 K	323 K	333 K	
-8.08	-7.03	-23.69	-23.30	-24.17	-25.01	0.8040

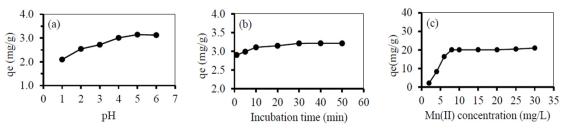


Fig. 1: Effects of (a) pH, (b) an incubation time and (c) an initial concentration of Mn(II)

Table 2: The kinetic parameters of Mn(II) adsorption using GO

<i>q_{e,exp}</i> (mg/g)	Pseudo-first order		Pseudo-secor	nd order	Intra-particle diffusion		
3.28	q _e (mg/g)	1.09	q _e (mg/g)	3.33	k _p (mg/g.min ^{1/2})	0.11	
	k ₁ (min)	0.15	k ₂ (g/mg.min)	0.38	C (mg/g)	2.65	
	R²	0.9936	R²	0.9999	R²	0.9915	

The effect of incubation time of Mn(II) adsorption onto the GO with 10 mg/L of the initial Mn(II) concentration at pH 5 is shown in Figure 1(b). It can be well seen that at the beginning the rate of adsorption went up with increasing incubation time with equilibrium time observed within 30 min. We, therefore, subjected our further experiments with an incubation time of 30 min. Figure 1(c) shows the equilibrium adsorption capacity plotted against the initial concentration of Mn(II). It is demonstrated that the equilibrium adsorption capacity increased with increasing initial concentration. At Mn(II) concentration of 8 mg/L, the maximum equilibrium adsorption capacity of 20.03 mg/g was reached. Higher collision probability between Mn(II) and the adsorbent particles together with yet-to-be-occupied active sites on the GO surface likely contributed to the initial increase in the uptake capacity of GO when an initial Mn(II) concentration was increased.

Thermodynamic study

For thermodynamic study, the calculations of Gibbs free energy change (ΔG^{0}), entropy change (ΔS^{0}) and enthalpy change (ΔH^{0}) were performed using the following equations:

$$\Delta G^{\circ} = -RT \ln K_d \qquad \dots (2)$$

$$Kd = \frac{q_e}{Ce} \qquad \dots (3)$$

$$\ln K_d = -\frac{\Delta H^\circ}{RT} + \frac{\Delta S^\circ}{R} \qquad \dots (4)$$

where, *R* is the constant (8.314 J/mol K), *T* is absolute temperature in K and K_d is the equilibrium constant. q_e is the amount of the Mn(II) at equilibrium (mg/g), C_e is the equilibrium concentration (mg/L) of Mn(II) in the solution.

The slope and the intercept of the van't Hoff plot of ln K_d versus 1/T, were used for the calculation of ΔS^0 and ΔH^0 , respectively (Fig. 2). Table 1 summarizes the thermodynamic data.

From Table 1, the negative value of ΔG^0 indicates that the ongoing adsorption process is spontaneous¹⁴. The negative ΔH^0 indicates that the adsorption of the metal by GO is exothermic in nature; while, the negative value of ΔS^0 indicates that during adsorption there is a tendency for lower disorder at the solid-solution interface¹⁵. The boundary layer becomes thinner at higher temperatures because metal molecules tend to escape from the adsorbent surface to its bulk solution, lowering the adsorption capacity when the temperature goes up¹⁶.

Kinetic study

To understand the nature of the adsorption phenomenon in terms of the order of the rate constant, three kinetic models were applied including pseudo-first order, pseudo-second-order and intraparticle diffusion model. These models are expressed in the following equations (Eq. (5) to Eq. (7)).

$$\log(q_e - q_t) = \log q_e - \frac{k_1}{2303}t \qquad \dots(5)$$

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e}t \qquad ...(6)$$

$$q_t = k_p t^{\frac{1}{2}} + C \qquad \dots (7)$$

where, k_i is rate constant (L/min) of pseudo-first order adsorption, k_2 is rate constant of pseudo-second-order adsorption (g/mg min), k_p is an intra-particle diffusion rate constant in mg/(g min^{1/2}) q_e and q_i are the amounts of Mn(II) adsorbed onto GO at an equilibrium state (mg/g) and at time t, respectively.

For the pseudo-first order, we calculated the values of k_1 and q_e from the slope and the intercept of the plot of log (q_e-q_t) versus t, respectively (Fig. 3(a)); while, the values of k_2 and q_e for the pseudo-second-order adsorption are calculated from the slope and the intercept of the plot of t/q_t versus t,

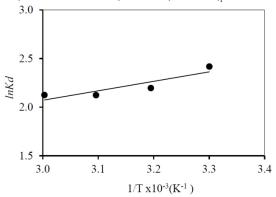


Fig. 2: Plot of In K_d versus 1/T for Mn(II) adsorption using GO

respectively (Fig. 3(b)). In addition, the values of k_{ρ} and *C* of the intra-particle diffusion were based on the calculation from the slope and the intercept of the plot of q_t versus $t^{1/2}$, respectively (Fig. 3(c)). The kinetic parameters obtained from these models are summarized in Table 2.

As shown in Table 2, the pseudo-first order kinetic model could not explain the adsorption of Mn(II) using GO, as there was a great difference between the value from the calculation of maximum adsorption capacity ($q_e = 1.09 \text{ mg/g}$) and that from the experiment ($q_{e,exp} = 3.28 \text{ mg/g}$). On the contrary, the calculated q_e value for the pseudo-second order kinetic model (3.33 mg/g) was similar to that from the experiment with high linear correlation coefficient (R^2) of 0.9999. Therefore, the pseudo-second-order kinetic model was applicable to the adsorption of Mn(II).

Adsorption isotherm

Langmuir and Freundlich models are used to describe the adsorption isotherms. The Langmuir model assumes that an adsorbate's monolayer coversa homogenous adsorbent surface which has a finite number of the adsorption sites via uniform strategies of adsorption without any transmigration of the adsorbate occurring along the surface plane. The Freundlich adsorption isotherm assumes a heterogeneous surface energy for which the energy term in the Langmuir equation varies as a function of a surface coverage. The linear form of both Langmuir and Freundlich isotherms are given by the following Eq. (8) and Eq. (9):

$$\frac{C_e}{q_e} = \frac{C_e}{q_m} + \frac{1}{q_m K_L} \tag{8}$$

$$\log q_e = \log K_F + \frac{1}{n} \log C_e \qquad \dots (9)$$

Table 3: Isotherm parameters and correlation coefficients calculated for the
adsorption of Mn(II) using GO

Adsorbent	Langmuir isotherm		Freundlich isotherm			
-	q _{max} (mg/g)	K	R ²	K _F	1/n	R ²
Graphene oxide	41.67	0.02	0.9765	2.99	0.88	0.8773

where, q_m is the maximum amount of Mn(II) absorbed per unit weight of adsorbent (mg/g) that forms complete monolayer on the surface at the equilibrium state of metal concentration (mg/L), q_e is the amount of Mn(II) adsorbed per unit weight of adsorbent at equilibrium, and K_L is the Langmuir constant (L/mg), K_F and n are the Freundlich constant and the intensity of adsorption, respectively.

The values of q_m and K_L are calculated from the slope and the intercept of the straight line plot of C_e/q_e versus C_e , respectively (Fig. 4(a)). The values of K_F and 1/n can be obtained from a linear plot of log q_e versus log C_e (Fig.4(b)). Table 3 shows the constant values that are obtained from the Langmuir and Freundlich adsorption isotherms and their correlation coefficients (R²)

From Table 3, it was found that Langmuir isotherm ($R^2 > 0.97$) fitted with the experimental results better than those of Freundlich isotherm ($R^2 > 0.87$). This indicates that there is a high degree of homogeneity on the GO surface with monolayer coverage of Mn(II) on it. The maximum adsorption capacity of Mn(II) was found to be 41.67 mg/g.

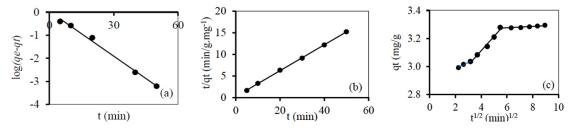


Fig. 3. (a) pseudo-first-order (b) pseudo-second-order and (c) intra-particle diffusion of Mn(II) adsorption using GO as an adsorbent

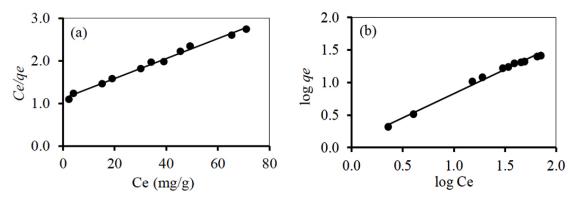


Fig. 4. (a) Langmuir and (b) Freundlich adsorption plots of Mn(II) adsorption using GO as an adsorbent

Table 4: Comparison of maximum adsorption capacities of M	In(II) using
various carbon-based adsorbents	

Adsorbent	q _m (mg/g)	Reference
Na-montmorillonite	3.22	[17]
Granular activated carbon	2.54	[18]
Dolomite	2.21	[19]
Activated carbon immobilized by tannic acid	1.73	[20]
Marble	1.20	[19]
Pithacelobium dulce carbon	0.41	[21]
Graphene oxide	41.67	This work

Table 4 shows the maximum adsorption capacity of Mn(II) using other adsorbents. In comparison with other carbon-based adsorbents, GO has by far the highest adsorption capacity for the Mn(II) removal from aqueous solution.

CONCLUSION

The results indicate the suitability and efficiency of graphene oxide as an adsorbent for the removal of Mn(II) from aqueous solution. With maximum adsorption capacity for Mn(II) being 41.67 mg/g, it far outstrips other carbon-based adsorbents. The pseudo second-order kinetic model was applicable to the adsorption kinetics.. The intraparticle diffusion model described that the intraparticle diffusion was not the only rate-limiting step, but surface adsorption and intra-particle diffusion mechanisms worked in simultaneous corporation during adsorption. ΔG^0 indicates the spontaneity of the ongoing adsorption process; while, ΔS^0 indicates a tendency to lower disorder at the solid-solution interface during the adsorption. In addition, the ΔH^0 indicates that the adsorption process is an exothermic nature. Our experiments clearly showed that the as-prepared GO has a high potential as adsorbent for Mn(II) removal.

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REFERENCES

- 1. Ates, A. *Powder Technol.* 2014, *264*, 86-95.
- 2. Flynn, MR.; Susi, P. *Int. J. Hyg. Environ Health.* 2009, *212*, 459-469.
- 3. Michalke, B.; Fernsebner, K. *J. Trace Elem. Med. Biol.* 2014, *28*, 106-116.
- Dusek, P.; Roos, P.M.; Litwin, T.; Schneider, S.A.; Flaten, T.P.; Aaseth, J. *J. Trace Elem. Med. Biol.* 2015, *31*, 193-203.
- 5. Nuengmatcha, P.; Mahachai, R.; Chanthai, S. Asian. J. Chem. 2014, 26, S85-S88.
- Madadrang, C.J; Kim, H.Y.; Gao, G.; Wang, N.; Zhu, J.; Feng, H.; Gorring, M.; Kasner, M.L.; Hou, S. ACS Appl. Mater. Interfaces. 2012, 4, 1186-1193.
- Zhang, K.; Dwivedi, V.; Chi, C.; Wu, J. J. Hazard. Mater. 2010, 182, 162-168.
- Deng, J.H.; Zhang, X.R.; Zeng, G.M.; Gong, J.L.; Niu, Q.Y.; Liang, J. *Chem. Eng. J.* 2013, *266*, 189-200.
- Li, L.; Fan, L.; Sun, M.; Qiu, H.; Li, X.; Duan, H.; Luo, C. *Colloids Surf.*, B 2013, 107, 76-83.
- Zhao, G.; Li, J.; Ren, X.; Chen, C.; Wang, X. *Environ. Sci. Technol.* 2011, *45*, 10454-10462.

- 11. Nuengmatcha, P.; Mahachai, R.; Chanthai, S. *Orient. J. Chem.* 2014, *30*, 1463-1474.
- 12. Nuengmatcha, P.; Mahachai, R.; Chanthai, S. *Orient. J. Chem.* 2016, *32*, 1399-1410.
- 13. Nuengmatcha, P.; Mahachai, R.; Chanthai, S. *Asian. J. Chem.*2014, *26*, 1321-1323.
- 14. Haghdoost, G. *Orient. J. Chem.* 2016, *32*(3), 1485-1492.
- Hadavifar, M.; Bahramifar, N.; Younesi, H.; Li, Q. *Chem. Eng. J.* 2014, *237*, 217-228.
- 16. Saha, P.; Chowdhury, S.; Gupta, S.; Kumar, I. *Chem. Eng. J.* 2010, *165*, 874-882.
- Abollino, O.; Aceto, M.; Malandrino, M.; Sarzanini, C.; Mentasti, E. *Water Res.* 2003, *37*, 1619-1627.
- Jusoh, A.B.; Cheng, W.H.; Low, W.M.; Nora'aini, A.; Noor, M.J.M.M. *Desalination*. 2005, *182*, 347-353.
- 19. Kroik, A.A.; Shramko, O.N.; Belous, N.V. *Chem. Technol. Water.* 1999, *21*, 310-314.
- 20. çer, A.—.; Uyanik, A.; Aygün, S.F. *Sep. Purif. Technol.* 2006, *47*, 113-118.
- 21. Emmanuela, K.A.; Veerabhadra Rao, A. *Rasayan J. Chem.* 2008, *1*, 840-852.